

# Resource partitioning along multiple niche axes drives functional diversity in parrotfishes on Caribbean coral reefs

Thomas C. Adam<sup>1,2</sup> · Megan Kelley<sup>1</sup> · Benjamin I. Ruttenberg<sup>3,4</sup> · Deron E. Burkepile<sup>1</sup>

Received: 22 February 2015 / Accepted: 16 July 2015 / Published online: 6 August 2015  
© Springer-Verlag Berlin Heidelberg 2015

**Abstract** The recent loss of key consumers to exploitation and habitat degradation has significantly altered community dynamics and ecosystem function across many ecosystems worldwide. Predicting the impacts of consumer losses requires knowing the level of functional diversity that exists within a consumer assemblage. In this study, we document functional diversity among nine species of parrotfishes on Caribbean coral reefs. Parrotfishes are key herbivores that facilitate the maintenance and recovery of coral-dominated reefs by controlling algae and provisioning space for the recruitment of corals. We observed large functional differences among two genera of parrotfishes that were driven by differences in diet. Fishes in the genus *Scarus* targeted filamentous algal turf assemblages, crustose coralline algae, and endolithic algae and avoided macroalgae, while fishes in the genus *Sparisoma* preferentially targeted macroalgae. However, species with similar diets

were dissimilar in other attributes, including the habitats they frequented, the types of substrate they fed from, and the spatial scale at which they foraged. These differences indicate that species that appear to be functionally redundant when looking at diet alone exhibit high levels of complementarity when we consider multiple functional traits. By identifying key functional differences among parrotfishes, we provide critical information needed to manage parrotfishes to enhance the resilience of coral-dominated reefs and reverse phase shifts on algal-dominated reefs throughout the wider Caribbean. Further, our study provides a framework for predicting the impacts of consumer losses in other species rich ecosystems.

**Keywords** Biodiversity · Fishing · Functional group · Herbivory · Niche partitioning

Communicated by Jeff Shima.

**Electronic supplementary material** The online version of this article (doi:10.1007/s00442-015-3406-3) contains supplementary material, which is available to authorized users.

✉ Thomas C. Adam  
adam@lifesci.ucsb.edu

<sup>1</sup> Marine Sciences Program, Department of Biological Sciences, Florida International University, North Miami, FL 33181, USA

<sup>2</sup> Present Address: Marine Science Institute, University of California, Santa Barbara, CA 93106, USA

<sup>3</sup> NOAA Fisheries, Southeast Fisheries Science Center, Miami, FL 33149, USA

<sup>4</sup> Biological Sciences Department, California Polytechnic State University, San Luis Obispo, CA 93410, USA

## Introduction

Ecosystems can undergo rapid transitions from a desirable state to a less desirable, degraded state with reduced capacity to provide important ecosystem services (e.g., regime shifts or phase shifts; Folke et al. 2004). These abrupt changes are frequently associated with the loss of key consumers that can lower the resilience of an ecosystem to natural and anthropogenic disturbances (Estes et al. 2011). However, in many systems, multiple consumers play similar roles, making it difficult to predict how the loss of any individual species will impact the overall structure and function of the ecosystem (Duffy 2002). This is especially true in species-rich systems such as tropical coral reefs and African savannahs, where dozens of consumers may perform similar ecological functions (Du Toit and Cumming 1999; Bellwood et al. 2004).

On coral reefs, herbivorous fishes and sea urchins are critical for maintaining ecosystem function by controlling algae that can displace reef-building corals. Herbivores influence reef ecosystems through a variety of processes, including the provisioning of space for coral settlement (Steneck et al. 2014) and the removal of macroalgae that can slow coral growth and increase coral mortality (Burkepile and Hay 2008). Loss of key herbivores can result in persistent phase shifts from coral-dominated systems to algal-dominated systems with reduced levels of primary and secondary production (Carpenter 1986) and compromised capacity to build reefs and provision habitat for other organisms (Perry et al. 2013; Bozec et al. 2015). Consequently, overexploitation of herbivores in subsistence and commercial fisheries could compromise the function of reef ecosystems by reducing the total biomass of herbivores and by altering the species composition and size structure of the herbivore assemblage (Edwards et al. 2014).

Diversity of herbivorous fishes can be important for maintaining ecosystem function on coral reefs because herbivores that feed on different types of algae have complementary impacts on benthic communities (Bellwood et al. 2006; Burkepile and Hay 2008; Rasher et al. 2013). Species that feed on filamentous algal turfs and associated detritus can facilitate coral recruitment and maintain reefs in a coral-dominated state, while species that feed on mature macroalgae can prevent macroalgae from overgrowing corals and can help reverse phase shifts on macroalgal-dominated reefs. The level of overlap in the diets of different herbivores (i.e., redundancy vs. complementarity), therefore, provides a metric of the level of functional diversity present within the herbivore guild (Burkepile and Hay 2011). Guilds with higher levels of redundancy among species may be more resilient to overharvesting and better able to fulfil their ecological function. However, herbivores vary widely in a number of traits in addition to diet that will modulate their impacts on reef ecosystems and influence patterns of redundancy, including size (Lokrantz et al. 2008), movement patterns (Nash et al. 2013), preferred habitats (Robertson and Gaines 1986), and the specific substrates they target while foraging (Brandl and Bellwood 2014). Coexistence theory predicts that species that are similar in one niche dimension, such as diet, will be different in others, such as habitat selection (MacArthur and Levins 1967). Thus, there may be less functional redundancy within the herbivore guild than predicted by patterns of diet overlap alone when one considers the many other dimensions of the niches of these species.

Over the last several decades, many coral reef ecosystems throughout the wider Caribbean have experienced a phase shift (Hughes et al. 2010). On many reefs, large, structurally complex corals have been replaced by fleshy algae and other non-reef building organisms, resulting

in the collapse of physical structure and the rapid loss of ecosystem function (Bruno et al. 2009; Hughes et al. 2010; Alvarez-Filip et al. 2011; Perry et al. 2013). The decline of Caribbean coral reefs has been linked to a variety of stressors such as climate change and compromised water quality, but the loss of herbivores due to overfishing and disease is likely a major contributor (Jackson et al. 2014). Although herbivore diversity in general appears important for facilitating healthy reefs (Burkepile and Hay 2008, 2010, 2011), we currently have a limited understanding of the level of functional diversity present within the herbivore guild (Adam et al. 2015a).

Here, we explored the level of functional redundancy versus complementarity that exists among herbivorous parrotfishes, the dominant grazers throughout much of the Caribbean (Mumby 2006). We used direct observation to document patterns of resource use for nine species of Caribbean parrotfishes. We then determined the extent that species attributes, such as phylogenetic history and body size, predict differences in foraging traits. We found that species that feed on similar types of algae often prefer different habitats, feed from different substrates, and forage at different spatial scales, indicating low levels of functional redundancy in the parrotfish guild. This information is crucial when managing parrotfishes for their capacity to prevent and reverse phase shifts to macroalgal dominance and facilitate corals.

## Materials and methods

### Organisms and study sites

Parrotfishes belong to two monophyletic clades, represented on Caribbean coral reefs by two genera, *Sparisoma* and *Scarus* (Streelman et al. 2002). *Sparisoma* parrotfishes are often associated with seagrass habitats in addition to reefs. They exhibit a high diversity of foraging modes, with some species feeding primarily on seagrasses and macroalgae while others predominantly scrape and excavate algae from carbonate substrates. In contrast, *Scarus* parrotfishes are almost exclusively reef-associated, with all species apparently adapted to scrape or excavate epilithic algal turfs and endolithic algae from carbonate substrates (Bonaldo et al. 2014). Species from both genera vary greatly in size (~2 orders of magnitude), and thus are likely to perform a range of different ecological functions.

This study was conducted during June and July 2013 in the Florida Keys National Marine Sanctuary (FKNMS) off of Key Largo, FL, USA. The Florida Keys Reef Tract consists of a large bank reef system located approximately 8 km offshore of the Florida Keys, USA, and paralleling the island chain. The FKNMS provided an ideal setting for

this study. Fishing pressure on parrotfishes in the Florida Keys is very low (Bohnsack et al. 1994). As a result there are large populations of several species of parrotfishes that are rare or absent in many other locations, including the largest parrotfishes in the Caribbean, *Scarus coelestinus*, *Scarus coeruleus*, and *Scarus guacamaia* (Paddack et al. 2006).

We conducted fish surveys on four shallow, high-relief spur and groove reefs (Molasses, French, Carysfort, and Elbow), with behavioral observations conducted on three of these (Molasses, Carysfort, and Elbow). Spur formations consist of large coral outcrops that rise approximately 2–6 m from the seafloor. Outcrops are interspersed with a mixture of sand, carbonate boulders, and small coral rubble, substrates which also dominate the shallow areas inshore of the primary spur and groove habitat. Offshore of the main spur formations, reefs transition to a slightly deeper low-relief carbonate platform (Fig. S1). We conducted fish surveys, behavioral observations, and benthic surveys in all three habitats: high-relief spur and groove (depth 2–6 m), low-relief carbonate platform/hardbottom (depth 4–12 m), and carbonate boulder/rubble fields (depth 4–9 m).

### Parrotfish surveys

At each site, we estimated parrotfish abundance in the three habitat types in order to assess the relative abundance and biomass of different species and to quantify differences in habitat selection. To estimate parrotfish density, we conducted 20–30 min timed swims while towing a GPS receiver on a float to measure the distance traveled and calculate the amount of area sampled. During a swim, the observer would swim parallel with the habitat type being sampled and count and estimate the size to the nearest cm of all parrotfishes  $\geq 15$  cm in length that were encountered in a 5-m-wide swath. The diver sampled multiple habitat types during each timed swim. Therefore, we divided the timed swims into 1-min intervals and recorded the habitat type during each interval (see Fig. S1b). The goal of the timed swims was to achieve similar sampling effort in each habitat type, and the method enabled us to sample  $\sim 30,000$  m<sup>2</sup> of habitat, an area sufficiently large to obtain density estimates of all parrotfishes including less common species (see Table S1 for amounts of each habitat type sampled). We calculated biomass of each species using published length-weight relationships (Bohnsack and Harper 1988).

### Behavioral observations

We characterized the diet and foraging behavior of the nine species of parrotfishes that account for >99 % of the

parrotfish biomass on these reefs, *Scarus coelestinus*, *Scarus coeruleus*, *Scarus guacamaia*, *Scarus taeniopterus*, *Scarus vetula*, *Sparisoma aurofrenatum*, *Sparisoma chrysopterum*, *Sparisoma rubripinne*, and *Sparisoma viride*. Approximately 18 individuals of each species (range = 16–19) were observed for 20 min each, with observations evenly distributed across three sites (see Table S2 for details). In order to control for diurnal variation in foraging behavior, we followed focal individuals in a balanced design in three 2-h sampling intervals (1000–1200, 1200–1400, and 1400–1600). This time period corresponds to peak feeding time for herbivorous fishes (e.g., Bruggemann et al. 1994). We focused on characterizing the behavior of large adults. For sexually dimorphic species, we limited observations to the initial phase (IP) because the IP individuals tend to be much more common than terminal phase individuals (Hawkins and Roberts 2003).

Focal individuals for behavioral observations were haphazardly selected using the following criteria. First, we planned our observations so that all species would be observed in the same general locations while also minimizing the potential for resampling the same individuals. We did this by following one individual of as many species as possible in a given location before moving to the next location. Second, we initiated our observations on or near high-relief habitat, which supported the highest density and biomass of most species. Third, we targeted the largest IP individuals we saw in these locations. After identifying a potential target individual, we approached the fish slowly and allowed  $\sim 2$ –3 min to acclimate to the presence of an observer. We estimated their size to the nearest cm and began observations. Fish usually acclimated quickly to the presence of an observer, but observations were occasionally aborted when fish did not resume normal foraging behavior. Foraging behavior was then recorded by a SCUBA diver for a period of 20 min while towing a GPS receiver (Garmin GPS 72) which obtained position fixes of the focal fish at 15-s intervals.

Fish were followed from a close distance ( $\sim 2$  m when possible), and food items were identified to the lowest taxonomic level possible, with macroalgae and coral usually identified to genus or species. Many bites involved scraping or excavating substrate colonized by a multi-species assemblage of filamentous “turf” algae, crustose coralline algae (CCA), and associated detritus, commonly referred to as the epilithic algal matrix (EAM) (Wilson et al. 2003). Because it was impossible to determine the specific food items targeted, these bites were categorized as being on EAM.

We also recorded the type of substrate targeted during each foraging bout, categorizing each substrate as one of the following: (1) dead coral, (2) coral pavement, (3) boulder, (4) rubble, or (5) ledge. Dead coral included both

convex and concave surfaces on the vertical and horizontal planes of three-dimensional coral skeletons (primarily dead *Acropora palmata*) that were attached to reef substrate. Coral pavement was carbonate reef with little topographic complexity (i.e., flat limestone pavement). Boulder was large remnants of dead mounding corals not clearly attached to the bottom and often partially buried in sand. Coral rubble consisted of small dead coral fragments (generally <10 cm in any dimension) that could be moved with minimal force. Ledges consisted entirely of the undercut sides of large spurs in spur and groove habitat. In addition to recording foraging behavior, we also recorded other activities such as aggressive interactions with other fish.

### Characterization of the benthos

To quantify the relative abundance of different food types, we estimated the percent cover of algae, coral, and other sessile invertebrates on each of the five substrates commonly targeted by parrotfishes (dead coral, coral pavement, boulder, rubble, or ledge) in 0.5 m × 0.5 m photoquadrats. We photographed a total of eight haphazardly selected quadrats dispersed throughout the study site for each substrate type at each of the three sites ( $N = 24$  quadrats per substrate type,  $N = 120$  quadrats total). Each photoquadrat was divided into sixteen 12 cm × 12 cm sections which were individually photographed, and percent cover was estimated from nine stratified random points per section ( $N = 144$  point per quadrat). Macroalgae and coral were identified to genus or species while other organisms were identified to functional group (e.g., sponges, gorgonians, turf algae, crustose coralline algae). For turf algae, we also determined whether a point landed on an algal filament or sediment bound in the turf.

### Questions and analyses

Our goals were to identify differences in foraging behavior and resource use among the nine species of parrotfish and to test the extent that each species exhibits a unique combination of foraging traits. To achieve this we: (1) investigated the degree that species differed in univariate traits (i.e., bite rates, distance travelled while foraging) and multivariate traits (i.e., diet, substrates grazed), (2) tested whether variation among species in foraging traits was related to size and/or genera, and (3) summarized relationships among species in multivariate trait-space.

We used linear mixed models to test for relationships between fish size (as estimated fork length), time of day, and genus on bite rate and the maximum linear distance moved during a 20-min observation. Models included species as a random effect and also tested for interactions between fish size and time of day and genus. Fork length, bite rate, and distance travelled were all log-transformed

prior to analyses to satisfy assumptions of the linear models. We then investigated the diet composition of each species using an electivity index (Manly's alpha; Chesson 1983) in order to identify whether particular food items tended to be eaten more or less often than expected based on their relative availability. We determined availability of food items by calculating mean abundance of each food type from point contacts of the major substrates targeted by parrotfishes. Electivity was calculated using weighted means of the percent cover of each food type on each substrate targeted by focal fish. Manly's alpha ranges from 0 to 1 and represents the predicted proportion of each food item included in an individual's diet if all food items were equally available. Differences in habitat selection were also evaluated by calculating an electivity index (Manly's alpha) based on the survey data. Next, differences in diet, substrates grazed, and habitat selection among species were visualized using hierarchical cluster analysis with Ward's linkage on Bray-Curtis dissimilarity matrices of square root transformed data. Differences among major clusters were then tested with permutational MANOVA (Anderson 2001) ( $N = \sim 18$  values per species). Permutation-based contingency tests (i.e., Fisher's exact) were then used to determine whether the different genera of parrotfish were distributed non-randomly among significant clusters.

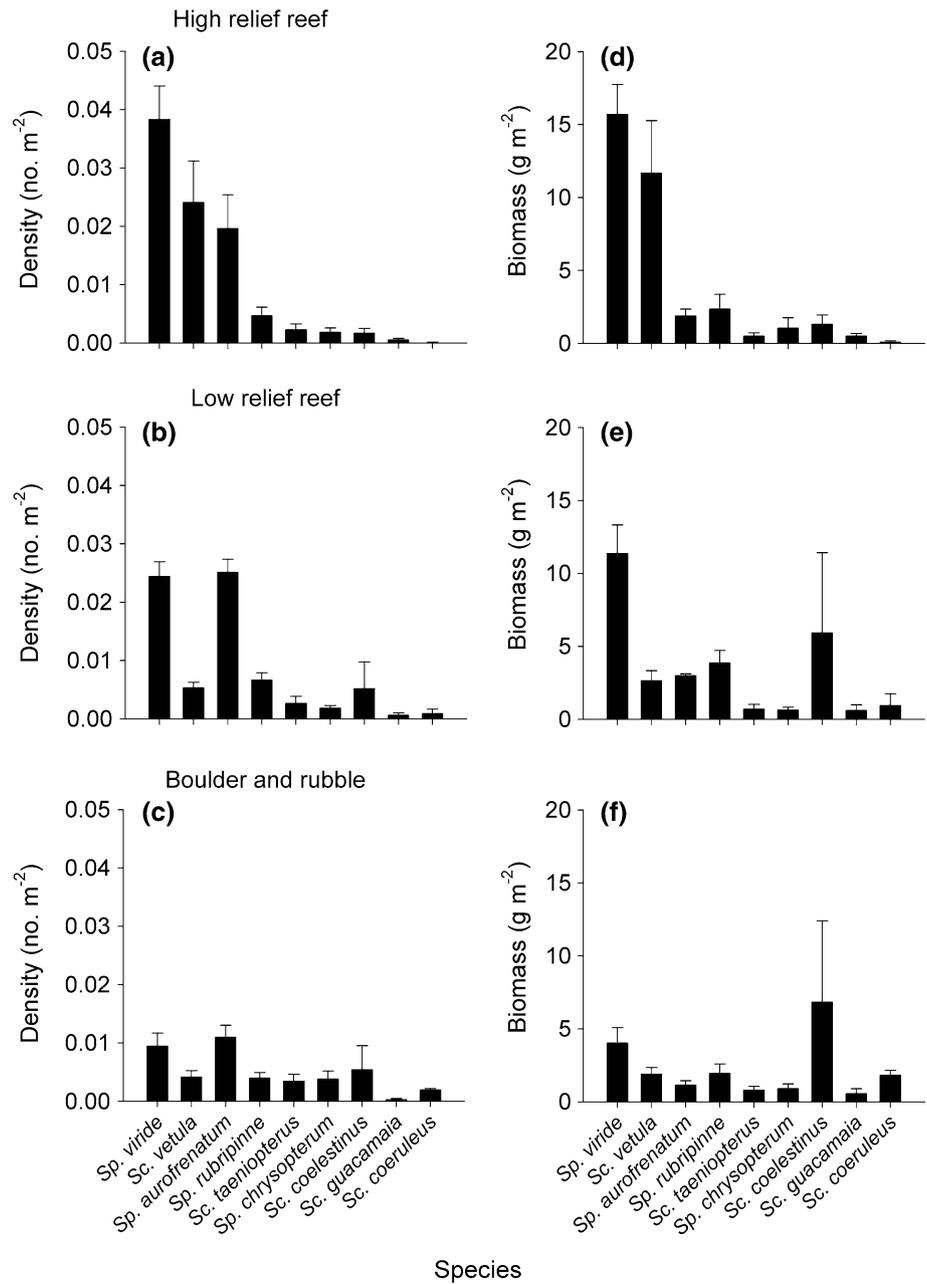
Finally, to summarize relationships among species in multivariate trait space (combining all traits such as diet, distance moved, substrate targeted, etc.), we first used three separate principal component analyses (PCA) to summarize data on diets, substrates targeted, and habitats selected, respectively, in one or two principal components (which explained at least 74 % of the variance for each resource type; Tables S3, S4, and S5). Next, principal components derived from the three analyses on each resource type were combined with data on bite rates and distance moved while foraging, normalized (mean zero and unit variance), and subjected to a redundancy analysis (RDA; Legendre and Legendre 2012), which ordinated species according to their combined trait values. Analyses were conducted in the R programming language using the vegan package (Oksanen et al. 2013) for multivariate analyses and the nlme (Pinheiro et al. 2013) and lmmfit (Maj 2011) packages for mixed effects models. All data are available from NOAA National Centers for Environmental Information: <http://accession.nodc.noaa.gov/0127525> (Adam et al. 2015b).

## Results

### Parrotfish community structure

Shallow, high-relief habitat on spur and groove reefs was dominated numerically by three species of parrotfishes,

**Fig. 1** Mean density and biomass (+SE) of nine species of parrotfish on (a, d) high relief reef, (b, e) low relief reef, and (c, f) boulder and rubble habitat (N = 4 sites)

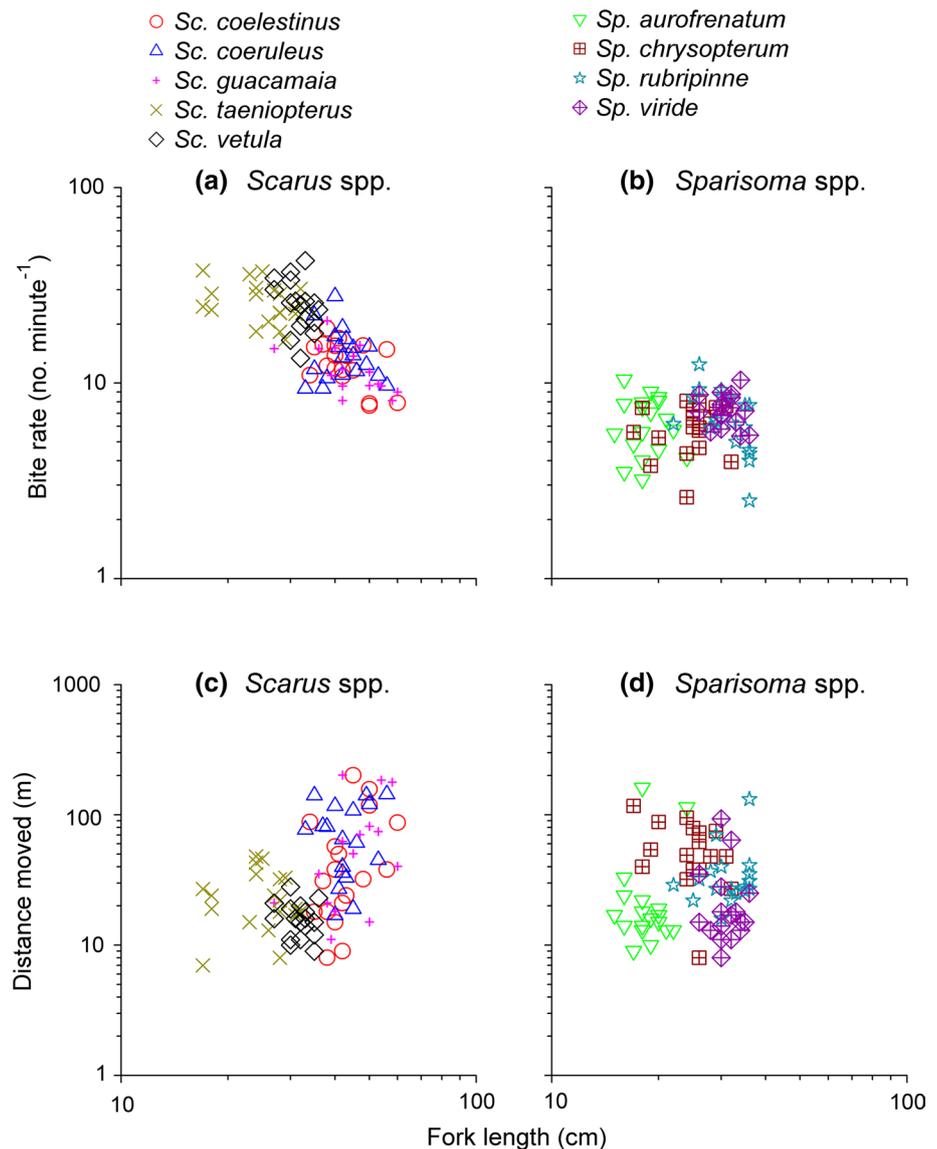


*Sp. viride*, *Sc. vetula*, and *Sp. aurofrenatum*, with low-relief hardbottom dominated by two of these (*Sp. viride* and *Sp. aurofrenatum*) (Fig. 1). *Sp. viride* and *Sp. aurofrenatum* were also the most abundant species in the boulder and rubble habitat, although this habitat had a more even mix of all nine species. *Sc. guacamaia* and *Sc. coeruleus* were the least abundant species in all habitats; however, both species contributed a non-trivial amount of biomass to the entire parrotfish assemblage due to their large size (2 and 3 %, respectively, compared to 7 % by numerically abundant but much smaller *Sp. aurofrenatum*).

**Diet and feeding ecology**

Both within and between genera, focal fishes varied greatly in size, bite rate, and distance moved (Fig. S2). There was a significant interaction between genus and fish length on bite rate (ANOVA,  $F_{1,149} = 7.40, P = 0.007$ ). Fish length and time of day both predicted bite rates of *Scarus* parrotfishes (full model  $R^2 = 0.52$ ), with bite rates decreasing with length (ANOVA,  $F_{1,82} = 21.8, P < 0.001$ ; Fig. 2a) and increasing throughout the day (ANOVA,  $F_{1,82} = 12.88, P < 0.001$ ). In contrast, bite rates of *Sparisoma* parrotfishes (full model  $R^2 = 0.08$ ) showed no correlation with

**Fig. 2** Relationship between length and bite rate for (a) *Scarus* and (b) *Sparisoma* parrotfishes, showing different relationships for each genus. Relationship between length and the maximum linear distance moved during an observation for (c) *Scarus* and (d) *Sparisoma* parrotfishes showing different relationships for each genus. Note the log scale on all graphs ( $N = \sim 18$  individuals per species; see Table S2 for exact sample sizes)

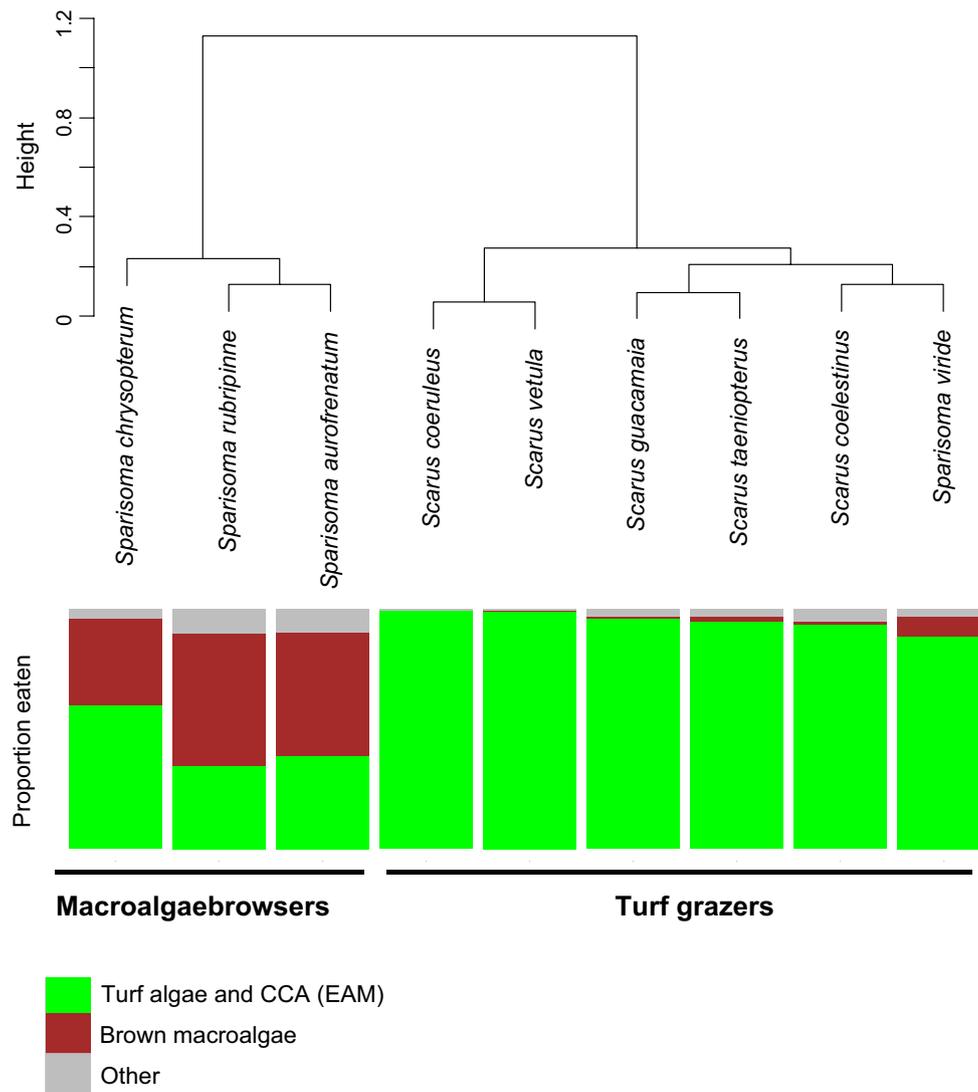


length (ANOVA,  $F_{1,67} = 0.042$ ,  $P = 0.84$ ; Fig. 2b), but a weak positive correlation with time of day (ANOVA,  $F_{1,67} = 6.09$ ,  $P = 0.016$ ).

There was also a significant interaction between genus and fish length on the distance moved during an observation (ANOVA,  $F_{1,147} = 5.41$ ,  $P = 0.021$ ). For *Scarus* parrotfishes, there was a significant positive correlation between length and distance moved (ANOVA,  $F_{1,80} = 11.76$ ,  $P = 0.001$ ; Fig. 2c), but no effect of time of day (ANOVA,  $F_{1,80} = 1.75$ ,  $P = 0.19$ ). For *Sparisoma* parrotfishes, there was no correlation between either length (ANOVA,  $F_{1,67} = 0.007$ ,  $P = 0.93$ ; Fig. 2d) or time of day (ANOVA,  $F_{1,67} = 0.683$ ,  $P = 0.41$ ) and distance moved.

EAM was the dominant food item on all substrate types, followed by brown macroalgae (primarily *Dictyota* spp.; Fig. S3a). Within the EAM complex, pavement, boulder, and rubble had high sediment levels with low cover of

CCA while sediment levels were lower and CCA higher on dead coral and ledges (Fig. S3). Parrotfishes fed primarily on the dominant food types (EAM and brown macroalgae). However, electivity indexes revealed that many parrotfishes also targeted less common types of macroalgae. For example, *Sp. chrysopterum* selected for red calcareous algae while *Sp. aurofrenatum* targeted green calcareous algae and *Sc. guacamaia* fed selectively on scleractinian corals and sponges (Fig. S4). Analysis of parrotfish diets revealed two distinct clusters of parrotfishes (PerMANOVA, Pseudo  $F_{1,160} = 230$ ,  $P < 0.001$ ; Fig. 3) with fish in the same genus tending to have similar diets (Fisher's exact test  $P = 0.048$ ). *Sparisoma viride* clustered with the *Scarus* parrotfishes, which fed primarily on EAM and endolithic algae. The three remaining species of *Sparisoma* formed a single cluster due to their tendency to feed on macroalgae, especially brown macroalgae (mainly *Dictyota* spp.).



**Fig. 3** Species-averaged dendrogram showing that parrotfishes cluster into two groups based on their diets. *Sp. chrysopteron*, *Sp. rubripinne*, and *Sp. aurofrenatum* feed largely on brown macroalgae (mainly *Dictyota* spp.), while fishes in the genus *Scarus* and

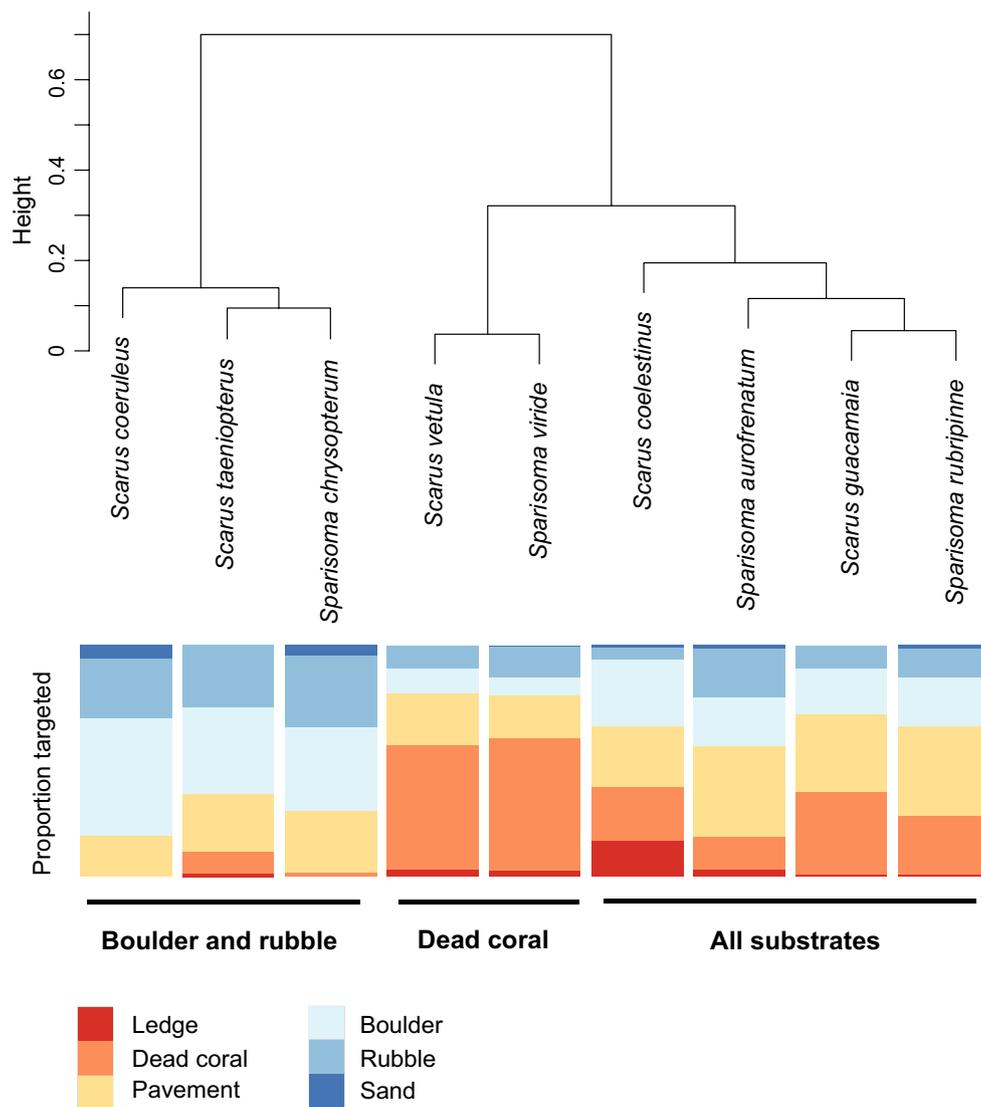
*Sp. viride* feed primarily on filamentous turfs, endolithic algae, and CCA. Bar charts show the mean proportion of each food item targeted ( $N = \sim 18$  individuals per species; see Table S2 for exact sample sizes). See Fig. S4 for less common food items and electivity values

While all species of parrotfishes targeted all major substrate types to some extent, multivariate analyses revealed three distinct clusters of species based on the substrates they targeted. The most distinct of these was a cluster consisting of *Sc. coeruleus*, *Sc. taeniopterus*, and *Sp. chrysopteron*, which primarily targeted carbonate boulders and coral rubble (Pseudo  $F_{1,160} = 38.7$ ,  $P < 0.001$ ; Fig. 4). Of the remaining six species, two clusters were evident, with *Sc. vetula* and *Sp. viride* primarily targeting dead coral substrate, and *Sc. coelestinus*, *Sp. aurofrenatum*, *Sc. guacamaia*, and *Sp. rubripinne* targeting a relatively equal proportion of all major substrate types (Pseudo  $F_{1,104} = 8.82$ ,  $P < 0.001$ ; Fig. 4). *Scarus* and *Sparisoma* parrotfishes were

randomly distributed across the three clusters (Fisher’s exact,  $P = 1$ ).

**Habitat selection**

While all parrotfishes were observed in all habitat types (Fig. 1), surveys revealed that different species exhibited distinct preferences for particular habitats (Fig. 5). Multivariate analyses revealed two distinct clusters of parrotfishes based on habitat preferences (Pseudo  $F_{1,34} = 10.51$ ,  $P < 0.001$ ; Fig. 5), with *Scarus* and *Sparisoma* parrotfishes randomly distributed across each (Fisher’s exact test,  $P = 0.357$ ). The first cluster, consisting of *Sc.*



**Fig. 4** Species-averaged dendrogram showing that parrotfishes cluster into three groups based on the substrates they target while foraging. *Sc. coeruleus*, *Sc. taeniopterus*, and *Sp. chrysopteron* primarily target carbonate boulder and coral rubble substrate. *Sc. vetula* and *Sp. vir-*

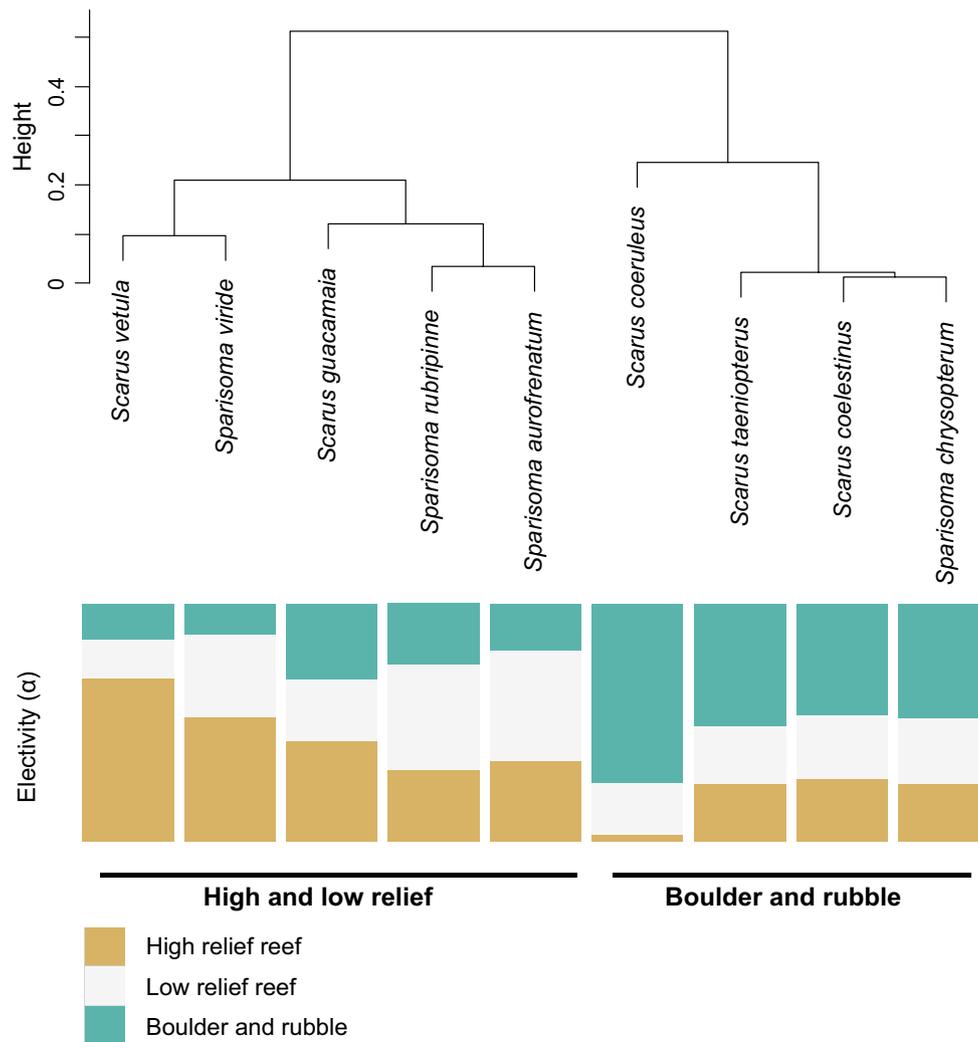
*ide* primarily target dead coral. *Sc. coelestinus*, *Sp. aurofrenatum*, *Sc. guacamaia*, and *Sp. rubripinne* target all substrates. Bar charts show the mean proportion of each substrate targeted ( $N = \sim 18$  individuals per species; see Table S2 for exact sample sizes)

*vetula*, *Sp. viride*, *Sc. guacamaia*, *Sp. aurofrenatum*, and *Sp. rubripinne* tended to avoid boulder and rubble habitat. The second cluster, consisting of *Sc. coeruleus*, *Sc. taeniopterus*, *Sc. coelestinus*, and *Sp. chrysopteron* tended to preferentially associate with boulder and rubble habitat (Fig. 5).

### Trait diversity

The RDA captured the majority of variance in traits in the first two axes, with these axes accounting for 47 and 38 % of the variance, respectively (Table S6). Differences in diet were mainly captured by RDA 2, with species that

feed primarily on turf algae having high values and species that feed primarily on macroalgae having low values (Fig. 6). In contrast, RDA 1 primarily captured differences in the spatial patterning of foraging, with species that travelled long distances and preferred boulder and rubble habitats having low values, and species that travelled short distances and preferred high relief habitat having high values (Fig. 6). While there was relatively little overlap between species that feed on macroalgae and those that primarily feed on EAM, there was overlap between some species within each of these groups. Among macroalgal browsers, there was extensive overlap between *Sp. aurofrenatum* and *Sp. rubripinne*; among turf grazers there was high overlap



**Fig. 5** Species-averaged dendrogram showing clustering of species based on electivity values for high relief, low relief, and boulder and rubble habitats from survey data. Species cluster into two significant groups, with *Sc. coeruleus*, *Sc. taeniopterus*, *Sc. coelestinus*, and *Sp.*

*chrysopterus* preferring boulder and rubble habitats, and all other species tending to avoid this habitat. Bar charts show mean electivity values for different habitat types ( $N = 4$  sites)

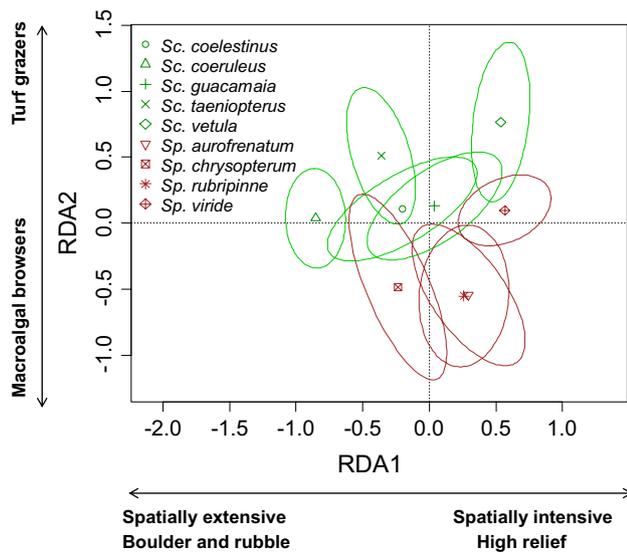
between *Sc. guacamaia* and *Sc. coelestinus*. Three of the turf grazers, *Sc. vetula*, *Sp. viride*, and *Sc. coeruleus* exhibited little overlap with other species (Fig. 6).

## Discussion

Herbivorous fishes can enhance the resilience of coral-dominated reefs by preventing the establishment and proliferation of algae that otherwise negatively impact coral settlement, growth, and survivorship (Hughes et al. 2007; Adam et al. 2011). Yet we are only beginning to understand the level of functional diversity present within the herbivore guild (Burkpile and Hay 2008; Hoey and Bellwood 2009; Rasher et al. 2013). In this study, we observed

large functional differences among herbivorous parrotfishes when we considered both diet and other important metrics like habitat selection and preferred feeding substrate. With one exception, parrotfishes belonging to the same genus fed on similar types of algae, suggesting that they may be functionally redundant in what they eat. However, species with similar diets were dissimilar in other attributes, such as the habitats they frequented or the types of substrate they fed from. These differences indicate that species that appear to be functionally redundant when looking at diet alone exhibit high levels of complementarity when we consider multiple functional traits that determine species' niches.

Previous work suggests that the diets of different genera of parrotfishes are often complementary. For example, Burkpile and Hay (2008, 2010) demonstrated



**Fig. 6** Ordination plot from redundancy analysis (RDA) of all foraging traits (i.e., bite rate, foraging range, diet, substrates targeted, and habitat preferences). Centroids for each species are plotted with dispersion ellipses using 0.9 confidence limits of the standard deviation of species scores. Fishes that feed primarily on EAM (turf grazers) have higher values of RDA 2 than those that feed on macroalgae. RDA 1 primarily captured differences in the spatial patterning of foraging, with spatially intensive foragers that prefer high relief habitat where they feed on dead coral having positive values, and spatially extensive foragers that prefer boulder and rubble habitat having negative values ( $N = 18$  individuals per species; see Table S2 for exact sample sizes)

experimentally that *Sp. aurofrenatum* and *Sc. taeniopterus* suppress algae more effectively together than either species can alone. However, these results were context-dependent with *Sc. taeniopterus* controlling filamentous algal turfs and preventing the establishment of macroalgae in early successional stage communities. Yet *Sp. aurofrenatum* prevented the spread and proliferation of macroalgae in later successional stage communities with high abundances of macroalgae, but had little impact on algal communities in early stages of succession. Our results indicate similar levels of diet complementarity among most species of *Sparisoma* and *Scarus* parrotfishes. The three species of *Sparisoma* parrotfishes that fed on significant amounts of brown macroalgae will be important for preventing macroalgae from encroaching and overgrowing corals and may help reverse phase shifts on macroalgal-dominated reefs. In contrast, *Scarus* parrotfishes and *Sp. viride*, which favor algal turfs, crustose coralline algae, and endolithic algae, can prevent the establishment of macroalgae and create bare space that can be colonized by coral larvae following a disturbance. Thus, a mix of species from each genus will likely be necessary to simultaneously prevent the spread of harmful algae and facilitate coral recruitment and growth.

When considered alone, our observations of diet among Caribbean parrotfishes suggest high levels of functional redundancy within each genus. However, in addition to knowing what different herbivores are eating, we also need to understand how their impacts are distributed in space. We found that species with similar diets often foraged in different locations on the reef. For example, both *Sc. vetula* and *Sc. coeruleus* targeted turf algae almost exclusively, but they foraged in different areas and targeted different substrates. *Sc. vetula* occupied small territories in high-relief areas where they fed predominantly on dead coral dominated by sparse turfs and CCA. But *Sc. coeruleus* roved over larger areas feeding on algal turfs with high sediment loads growing on loose coral rubble and carbonate boulders. Similarly, *Sp. chrysopterus* and *Sp. rubripinne* both fed on significant amounts of brown macroalgae, but *Sp. rubripinne* frequently fed from dead coral in high relief areas while *Sp. chrysopterus* primarily fed from unconsolidated coral rubble and carbonate boulders. Closely related herbivores with similar diets and foraging modes may, therefore, have fundamentally different impacts on reef ecosystems due to different habitat preferences and preferred feeding substrate.

Variation in where herbivores forage is likely to have important consequences for the maintenance and recovery of coral-dominated reefs since some locations within a reef will be more suitable for the growth and recruitment of corals and algae than others. Species that feed intensely in high relief areas, such as *Sc. vetula* and *Sp. viride*, may be especially important for creating and maintaining habitat suitable for coral recruitment. In contrast, species that preferentially associate with boulder and rubble habitat may be functionally less important since they tend to feed on algae growing on unstable coral rubble and carbonate boulders with high sediment loads which are poor habitat for corals (Birrell et al. 2005).

Differences in the movement behavior of herbivores can also impact the spatial patterning of primary producer communities, a phenomenon that has been well studied in grassland systems (Bakker et al. 1984; Adler et al. 2001; de Knegt et al. 2008). For example, on the African savannah, grazing by ungulates can stimulate productivity of grasses. As a result, grazers frequently revisit previously grazed patches, thereby creating mosaics of short, high-quality grass interspersed with lightly grazed, low-quality patches (McNaughton 1984). Similarly, in California grasslands, intense grazing by pocket gophers facilitates the long-term persistence of annual plants within their territories, while in the absence of grazing, the remaining landscape becomes dominated by late successional perennial species that deter colonization by gophers (Seabloom and Richards 2003). In both cases, positive feedback between grazing and forage

quality interact with herbivore behavior to create persistent landscape scale patterns of heterogeneity in plant communities. Similar dynamics are likely in play on coral reefs, where intense grazing by territorial species such as *Sc. vetula* may help maintain algal communities in a highly palatable state dominated by fast-growing filamentous turf algae with high cover of CCA which can facilitate coral recruitment. In contrast, grazing by more wide-ranging fishes, such as *Sc. coeruleus*, is less likely to create the same positive feedbacks unless overall grazing levels are very high, or fishes consistently return to the same patches to forage (Sandin and McNamara 2012).

The high levels of complementarity and low levels of redundancy we observed among Caribbean parrotfishes suggests that loss of any single species is likely to result in the loss of some function. However, predicting the impact of species losses on ecosystem function requires understanding the degree of plasticity in a species' functional traits. If plasticity is high and species adjust their traits in response to competitors, competition could lead to underestimates of a species functional niche. For example, complementarity in habitat use could be driven by interference competition, with competitively dominant species excluding subordinate species from the highest quality habitats (e.g., Robertson and Gaines 1986). We observed many aggressive interactions between parrotfishes (146 in 162 focal observations), suggesting that competition was intense and that habitat use may be influenced by competitive interactions. Most aggressive interactions (95 %) occurred between species in the same genus (which had largely overlapping diets), and the majority (60 %) were interspecific, suggesting that competition for food may drive closely related species to forage in different habitats and target different substrates. Many of the aggressive interactions we observed involved *Sc. vetula* aggressively defending their small territories in high-relief habitat. Substrate in the high-relief habitat was dominated by sparse turfs with low levels of sediments and macroalgae, a high quality diet relative to more highly sedimented turfs or dense macroalgae that often grew on coral rubble and low-relief hardbottom (Reinthal and Lewis 1986; McClanahan et al. 2000; Bellwood and Fulton 2008). Thus, it appears likely that *Sc. vetula* excludes other turf grazers from high-relief habitat, and the loss of *Sc. vetula* could result in the expansion of the functional niches of its competitors. Alternatively, the loss of *Sc. vetula* could result in a reduction in grazing intensity in these high-relief areas if other species only partially compensated for the high levels of herbivory by *Sc. vetula* in these habitats. Future work aimed at understanding how competition alters diet and habitat use would be especially useful for understanding the amount of redundancy present within the parrotfish guild.

Patterns of functional redundancy and complementarity could also vary with benthic community composition. Indeed, while our observations of diet are broadly consistent with previous examinations of subsets of Caribbean parrotfishes in other locations (Randall 1967; Lewis 1985; Lewis and Wainwright 1985; Bruggemann et al. 1994; McAfee and Morgan 1996; Cardoso et al. 2009; Burkepile and Hay 2010), the particular types of algae targeted by species can vary greatly among systems. For example, Burkepile and Hay (2011) found that large *Scarus* species, including *Sc. coelestinus*, *Sc. guacamaia*, and *Sc. vetula* all apparently fed preferentially on articulated red coralline algae that grew in experimental herbivore enclosures. These observations suggest that large *Scarus* species may be more similar to *Sparisoma* species than they are to smaller *Scarus* species (such as *Sc. taeniopterus*) with regards to their strong preferences for some types of erect calcified macroalgae. Work on large *Scarus* parrotfishes in the Indo-Pacific also indicate that these 'scraping and excavating grazers' frequently preferentially browse on some types of erect calcified red and green algae when given the chance (Mantyka and Bellwood 2007; Rasher et al. 2013; Hamilton et al. 2014). These data suggest that the foraging patterns of some species may be quite flexible and that studies using behavioral observations to document feeding only on the common algal species may overestimate similarity in the diet preferences among species. Our electivity analyses support these ideas as several parrotfishes targeted relatively rare algal taxa such as articulated red coralline algae. Understanding which species control these rarer algal taxa is important as these typically rare species are often the taxa that come to dominate areas when levels of herbivory are reduced (e.g., Burkepile and Hay 2008). Thus, more detailed work on how diet preferences change across varying levels of algal community composition are needed to help resolve these relationships.

Scientists and managers increasingly recognize the importance of functional diversity for maintaining healthy resilient ecosystems (Peterson et al. 1998; Bellwood et al. 2004; Cadotte et al. 2011). Species are being lost from many ecosystems at an alarming rate, with large consumers often the first to go (Dirzo et al. 2014; McCauley et al. 2015). Predicting the impact of species losses on ecosystem function, therefore, requires knowing how much functional diversity exists within consumer assemblages. We found high levels of functional diversity among Caribbean parrotfishes, key herbivores that facilitate reef-building corals. While closely related species had similar diets, they preferred different habitats, fed from different substrates, and foraged at different spatial scales. Subtle differences in habitat selection and the spatial scale of foraging will strongly modify feedbacks

between herbivores, algae, and corals, emphasizing the need to consider multiple functional traits when predicting the impact of species losses in high diversity systems. By identifying key functional differences among Caribbean parrotfishes, this study provides critical information needed to manage parrotfishes to enhance the resilience of coral-dominated reefs and reverse phase shifts on algal-dominated reefs throughout the wider Caribbean.

**Acknowledgments** This work was supported by a grant from the NOAA Coral Reef Conservation Program to BIR and DEB and facilitated by grant number OCE-1130786 from the National Science Foundation to DEB and R. Vega Thurber. We thank Laura Catano for assistance with fieldwork and Andrew Hoey and one anonymous reviewer for comments that improved the manuscript. This work was conducted with permission from the Florida Keys National Marine Sanctuary under permit no. FKNMS-2013-058. The authors have no conflicts of interest to declare.

**Author contribution statement** TCA designed the study with input from DEB and BIR. TCA and MK conducted field work. TCA developed methodology and performed analyses. TCA and DEB wrote the manuscript; BIR provided editorial advice.

## References

- Adam TC, Schmitt RJ, Holbrook SJ et al (2011) Herbivory, connectivity, and ecosystem resilience: response of a coral reef to a large-scale perturbation. *PLoS One* 6:e23717. doi:10.1371/journal.pone.0023717
- Adam TC, Burkepile DE, Ruttenberg BI, Paddock MJ (2015a) Herbivory and the resilience of Caribbean coral reefs: knowledge gaps and implications for management. *Mar Ecol Prog Ser* 520:1–20. doi:10.3354/meps11170
- Adam TC, Kelley M, Ruttenberg BI, Burkepile DE (2015) Abundance and behavior of parrotfishes (Labridae, Scarinae) in the upper Florida Keys from 2013-06-19 to 2013-07-30 (NODC Accession 0127525)
- Adler PB, Raff DA, Lauenroth WK (2001) The effect of grazing on the spatial heterogeneity of vegetation. *Oecologia* 128:465–479. doi:10.1007/s004420100737
- Alvarez-Filip L, Côté IM, Gill JA et al (2011) Region-wide temporal and spatial variation in Caribbean reef architecture: is coral cover the whole story? *Glob Chang Biol* 17:2470–2477. doi:10.1111/j.1365-2486.2010.02385.x
- Anderson MJ (2001) A new method for non-parametric multivariate analysis of variance. *Austral Ecol* 26:32–46
- Bakker J, de Leeuw J, van Wieren S (1984) Micro-patterns in grassland vegetation created and sustained by sheep-grazing. *Plant Ecol* 55:153–161
- Bellwood DR, Fulton CJ (2008) Sediment-mediated suppression of herbivory on coral reefs: decreasing resilience to rising sea-levels and climate change? *Limnol Oceanogr* 53:2695–2701. doi:10.4319/lo.2008.53.6.2695
- Bellwood DR, Hughes TP, Folke C, Nyström M (2004) Confronting the coral reef crisis. *Nature* 429:827–833. doi:10.1038/nature02691
- Bellwood DR, Hughes TP, Hoey AS (2006) Sleeping functional group drives coral-reef recovery. *Curr Biol* 16:2434–2439. doi:10.1016/j.cub.2006.10.030
- Birrell CL, McCook LJ, Willis BL (2005) Effects of algal turfs and sediment on coral settlement. *Mar Pollut Bull* 51:408–414. doi:10.1016/j.marpolbul.2004.10.022
- Bohnsack JA, Harper DE (1988) Length-weight relationships of selected marine reef fishes from the southeastern United States and the Caribbean. NOAA technical memorandum NMFS-SEFC-215, pp 1–31
- Bohnsack JA, Harper DE, McClellan DB (1994) Fisheries trends from Monroe County, Florida. *Bull Mar Sci* 54:982–1018
- Bonaldo RM, Hoey AS, Bellwood DR (2014) The ecosystem roles of parrotfishes on tropical reefs. *Oceanogr Mar Biol Annu Rev* 52:81–132
- Bozec Y-M, Alvarez-Filip L, Mumby PJ (2015) The dynamics of architectural complexity on coral reefs under climate change. *Glob Chang Biol* 21:223–235. doi:10.1111/gcb.12698
- Brandl SJ, Bellwood DR (2014) Individual-based analyses reveal limited functional overlap in a coral reef fish community. *J Anim Ecol* 83:661–670. doi:10.1111/1365-2656.12171
- Bruggemann JH, Kuyper MWM, Breeman AM (1994) Comparative analysis of foraging and habitat use by the sympatric Caribbean parrotfish *Scarus vetula* and *Sparisoma viride* (Scaridae). *Mar Ecol Prog Ser* 112:51–66
- Bruno JF, Sweatman H, Precht WF (2009) Assessing evidence of phase shifts from coral to macroalgal dominance on coral reefs. *Ecology* 90:1478–1484
- Burkepile DE, Hay ME (2008) Herbivore species richness and feeding complementarity affect community structure and function on a coral reef. *Proc Natl Acad Sci USA* 105:16201–16206. doi:10.1073/pnas.0801946105
- Burkepile DE, Hay ME (2010) Impact of herbivore identity on algal succession and coral growth on a Caribbean reef. *PLoS One* 5:e8963. doi:10.1371/journal.pone.0008963
- Burkepile DE, Hay ME (2011) Feeding complementarity versus redundancy among herbivorous fishes on a Caribbean reef. *Coral Reefs* 30:351–362. doi:10.1007/s00338-011-0726-6
- Cadotte MW, Carscadden K, Mirotchnick N (2011) Beyond species: functional diversity and the maintenance of ecological processes and services. *J Appl Ecol* 48:1079–1087. doi:10.1111/j.1365-2664.2011.02048.x
- Cardoso SC, Soares MC, Oxenford HA, Côté IM (2009) Interspecific differences in foraging behaviour and functional role of Caribbean parrotfish. *Mar Biodivers Rec* 2:e148. doi:10.1017/S1755267209990662
- Carpenter RC (1986) Partitioning herbivory and its effects on coral reef algal communities. *Ecol Monogr* 56:345–364
- Chesson J (1983) The estimation and analysis of preference and its relationship to foraging models. *Ecology* 64:1297–1304
- De Knegt HJ, Groen TA, van de Vijver CADM et al (2008) Herbivores as architects of savannas: inducing and modifying spatial vegetation patterning. *Oikos*. doi:10.1111/j.2008.0030-1299.16403.x
- Dirzo R, Young H, Galetti M, Ceballos G (2014) Defaunation in the Anthropocene. *Science* 345:401–406
- Du Toit JT, Cumming DHM (1999) Functional significance of ungulate diversity in African savannas and the ecological implications of the spread of pastoralism. *Biodivers Conserv* 8:1643–1661
- Duffy JE (2002) Biodiversity and ecosystem function: the consumer connection. *Oikos* 99:201–219. doi:10.1034/j.1600-0706.2002.990201.x
- Edwards CB, Friedlander AM, Green AG et al (2014) Global assessment of the status of coral reef herbivorous fishes: evidence for fishing effects. *Proc R Soc B Biol Sci* 281:20131835. doi:10.1098/rspb.2013.1835
- Estes JA, Terborgh J, Brashares JS et al (2011) Trophic downgrading of planet Earth. *Science* 333:301–306. doi:10.1126/science.1205106
- Folke C, Carpenter S, Walker B (2004) Regime shifts, resilience, and biodiversity in ecosystem management. *Annu Rev Ecol Evol Syst* 35:557–581. doi:10.2307/annurev.ecolsys.35.021103.30000021
- Hamilton SL, Smith JE, Price NN, Sandin SA (2014) Quantifying patterns of fish herbivory on Palmyra Atoll (USA), an uninhabited

- predator-dominated central Pacific coral reef. *Mar Ecol Prog Ser* 501:141–155. doi:[10.3354/meps10684](https://doi.org/10.3354/meps10684)
- Hawkins JP, Roberts CM (2003) Effects of fishing on sex-changing Caribbean parrotfishes. *Biol Conserv* 115:213–226. doi:[10.1016/S0006-3207\(03\)00119-8](https://doi.org/10.1016/S0006-3207(03)00119-8)
- Hoey AS, Bellwood DR (2009) Limited functional redundancy in a high diversity system: single species dominates key ecological process on coral reefs. *Ecosystems* 12:1316–1328. doi:[10.1007/s10021-009-9291-z](https://doi.org/10.1007/s10021-009-9291-z)
- Hughes TP, Rodrigues MJ, Bellwood DR et al (2007) Phase shifts, herbivory, and the resilience of coral reefs to climate change. *Curr Biol* 17:360–365. doi:[10.1016/j.cub.2006.12.049](https://doi.org/10.1016/j.cub.2006.12.049)
- Hughes TP, Graham NAJ, Jackson JBC et al (2010) Rising to the challenge of sustaining coral reef resilience. *Trends Ecol Evol* 25:633–642. doi:[10.1016/j.tree.2010.07.011](https://doi.org/10.1016/j.tree.2010.07.011)
- Jackson J, Donovan M, Cramer K, Lam V (eds) (2014) Status and Trends of Caribbean Coral Reefs 1970–2012. Global Coral Reef Monitoring Network c/o International Union for the Conservation of Nature Global Marine and Polar Program, Washington, D. C
- Legendre P, Legendre L (2012) Numerical ecology. Third English Edition. Elsevier, Amsterdam
- Lewis SM (1985) Herbivory on coral reefs: algal susceptibility to herbivorous fishes. *Oecologia* 65:370–375
- Lewis SM, Wainwright PC (1985) Herbivore abundance and grazing intensity on a Caribbean coral reef. *J Exp Mar Bio Ecol* 81:215–228
- Lokrantz J, Nyström M, Thyresson M, Johansson C (2008) The non-linear relationship between body size and function in parrotfishes. *Coral Reefs* 27:967–974. doi:[10.1007/s00338-008-0394-3](https://doi.org/10.1007/s00338-008-0394-3)
- MacArthur R, Levins R (1967) The limiting similarity, convergence, and divergence of coexisting species. *Am Nat* 101:377–385
- Maj A (2011) lmmfit: goodness-of-fit measures for linear mixed models with one-level-grouping. R package version 1.0
- Mantyka CS, Bellwood DR (2007) Macroalgal grazing selectivity among herbivorous coral reef fishes. *Mar Ecol Prog Ser* 352:177–185. doi:[10.3354/meps07055](https://doi.org/10.3354/meps07055)
- McAfee ST, Morgan SG (1996) Resource use by five sympatric parrotfishes in the San Blas Archipelago, Panama. *Mar Biol* 125:427–437
- McCauley DJ, Pinsky ML, Palumbi SR et al (2015) Marine defaunation: animal loss in the global ocean. *Science* 347:1255641. doi:[10.1126/science.1255641](https://doi.org/10.1126/science.1255641)
- McClanahan TR, Bergman K, Huitric M et al (2000) Response of fishes to algae reduction on Glovers Reef, Belize. *Mar Ecol Prog Ser* 206:273–282. doi:[10.3354/meps206273](https://doi.org/10.3354/meps206273)
- McNaughton S (1984) Grazing lawns: animals in herds, plant form, and coevolution. *Am Nat* 124:863–886
- Mumby PJ (2006) The impact of exploiting grazers (Scaridae) on the dynamics of Caribbean coral reefs. *Ecol Appl* 16:747–769
- Nash KL, Graham N, Bellwood DR (2013) Fish foraging patterns, vulnerability to fishing and implications for the management of ecosystem function across scales. *Ecol Appl* 23:1632–1644
- Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara RB, Simpson GL, Solymos P, Stevens MHH, Wagner H (2013) vegan: community ecology package. R package version 2.0–10
- Paddack MJ, Cowen RK, Sponaugle S (2006) Grazing pressure of herbivorous coral reef fishes on low coral-cover reefs. *Coral Reefs* 25:461–472. doi:[10.1007/s00338-006-0112-y](https://doi.org/10.1007/s00338-006-0112-y)
- Perry CT, Murphy GN, Kench PS et al (2013) Caribbean-wide decline in carbonate production threatens coral reef growth. *Nat Commun* 4:1402. doi:[10.1038/ncomms2409](https://doi.org/10.1038/ncomms2409)
- Peterson G, Allen C, Holling C (1998) Ecological resilience, biodiversity, and scale. *Ecosystems* 1:6–18
- Pinheiro J, Bates D, DebRoy S, Sarkar D; R Development Core Team (2013). nlme: linear and nonlinear mixed effects models. R package version 3.1-113
- Randall JE (1967) Food habits of reef fishes of the West Indies. *Stud Trop Oceanogr* 5:665–847
- Rasher DB, Hoey AS, Hay ME (2013) Consumer diversity interacts with prey defenses to drive ecosystem function. *Ecology* 94:1347–1358
- Reinthal P, Lewis S (1986) Social behaviour, foraging efficiency and habitat utilization in a group of tropical herbivorous fish. *Anim Behav* 34:1687–1693
- Robertson DR, Gaines SD (1986) Interference competition structures habitat use in a local assemblage of coral reef surgeonfishes. *Ecology* 67:1372–1383
- Sandin SA, McNamara DE (2012) Spatial dynamics of benthic competition on coral reefs. *Oecologia* 168:1079–1090. doi:[10.1007/s00442-011-2156-0](https://doi.org/10.1007/s00442-011-2156-0)
- Seabloom E, Richards S (2003) Multiple stable equilibria in grasslands mediated by herbivore population dynamics and foraging behavior. *Ecology* 84:2891–2904
- Steneck RS, Arnold SN, Mumby PJ (2014) Experiment mimics fishing on parrotfish: insights on coral reef recovery and alternative attractors. *Mar Ecol Prog Ser* 506:115–127. doi:[10.3354/meps10764](https://doi.org/10.3354/meps10764)
- Streelman JT, Alfaro M, Westneat MW (2002) Evolutionary history of the parrotfishes: biogeography, ecomorphology, and comparative diversity. *Evolution (N Y)* 56:961–971
- Wilson SK, Bellwood DR, Choat JH, Furnas MJ (2003) Detritus in the epilithic algal matrix and its use by coral reef fishes. *Oceanogr Mar Biol* 41(41):279–309

## **ELECTRONIC SUPPLEMENTARY MATERIAL (ESM)**

### **Resource partitioning along multiple niche axes drives functional diversity in parrotfishes on Caribbean coral reefs**

Thomas C. Adam\*<sup>1,2</sup>, Megan Kelley<sup>1</sup>, Benjamin I. Ruttenberg<sup>3</sup>, and Deron E. Burkepile<sup>1</sup>,

<sup>1</sup>Marine Sciences Program, Department of Biological Sciences, Florida International University,  
North Miami, FL, USA 33181

<sup>2</sup> *Present Address*: Marine Science Institute, University of California, Santa Barbara, CA USA  
93106

<sup>3</sup>NOAA Fisheries, Southeast Fisheries Science Center, Miami, FL USA 33149 and Biological  
Sciences Department, California Polytechnic State University, San Luis Obispo, CA USA  
93410

\*thomas.adam@lifesci.ucsb.edu

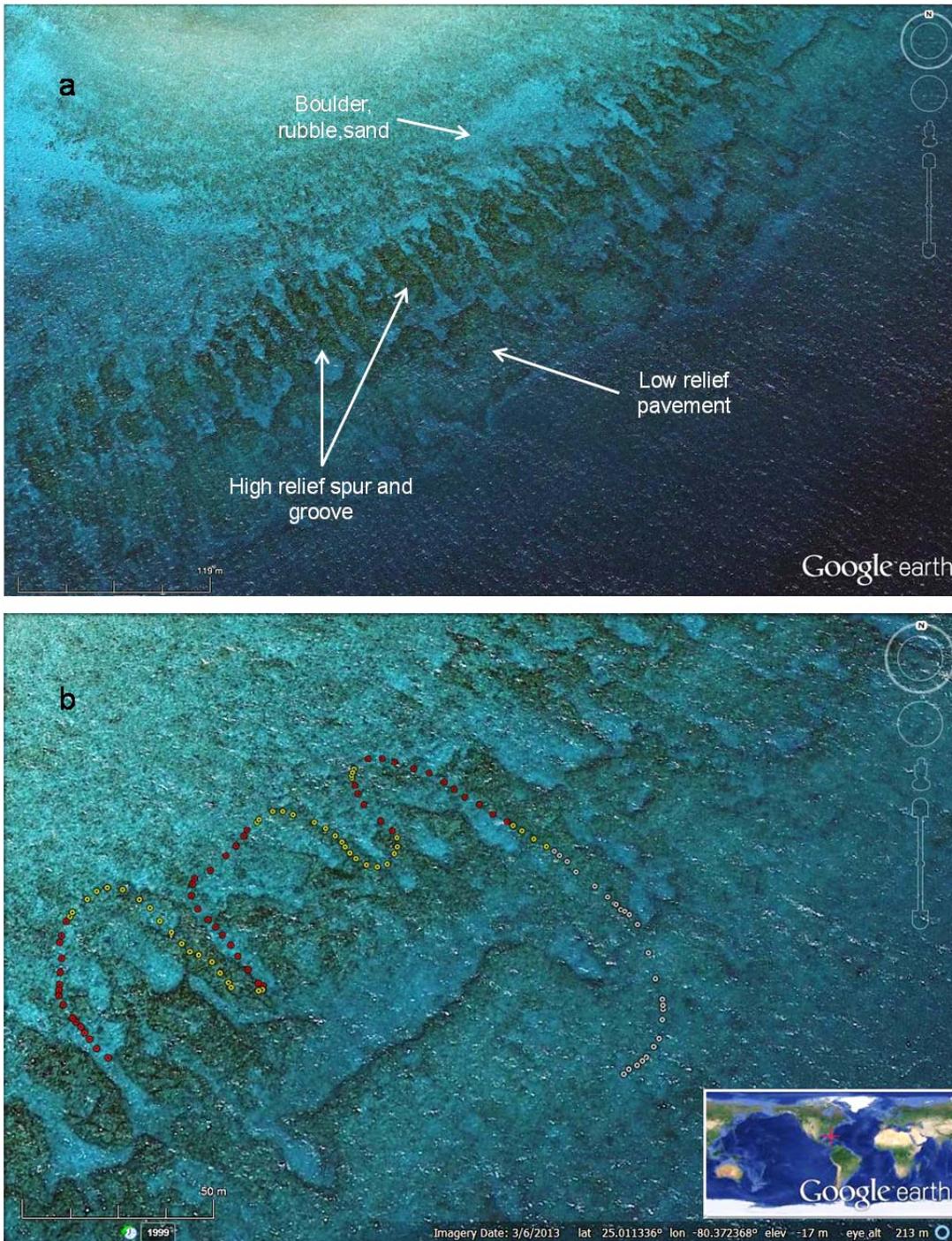


Figure S1. (a) Satellite image of Molasses reef ( $25.01^{\circ}$ ,  $-80.37^{\circ}$ ) off of Key Largo, FL, USA showing major habitat types occupied by parrotfishes. Scale bar in lower left corner is 119 m (b) Track from a 30 minute timed swim at Molasses reef where all three habitat types were sampled. Waypoints were taken every 15 s and are shown by colored dots (red = boulder/rubble habitat, Yellow = high relief reef, grey = low-relief carbonate platform/hardbottom). Scale bar in lower left corner is 50 m.

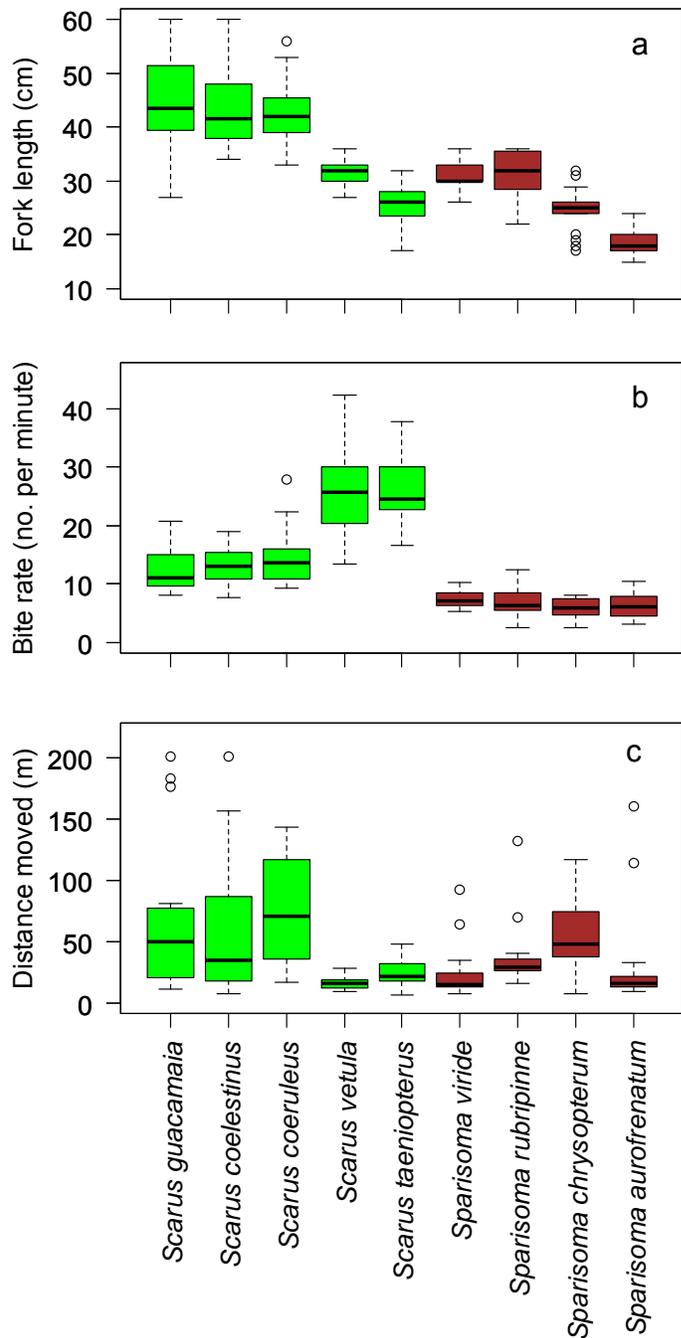


Figure S2. Box and whisker plots showing the (a) estimated size, (b) bite rate, and (c) distance moved of focal fish during 20 min behavioral observations (N = ~ 18 per species; see table S2 for exact sample sizes). Boxes for *Scarus* spp. are colored green and those for *Sparisoma* spp. are brown. The two smaller species of *Scarus*, *Sc. vetula* and *Sc. taeniopterus* had the highest bite rates, while species in the genus *Sparisoma* had the lowest bite rates, with larger *Scarus* spp. being intermediate between the two groups. The three largest species of *Scarus*, *Sc. guacamaia*, *Sc. coelestinus*, and *Sc. coeruleus*, together with *Sp. chrysopterus* tended to have the largest foraging ranges, while *Sc. vetula*, *Sp. viride*, and *Sp. aurofrenatum* had the smallest ranges.

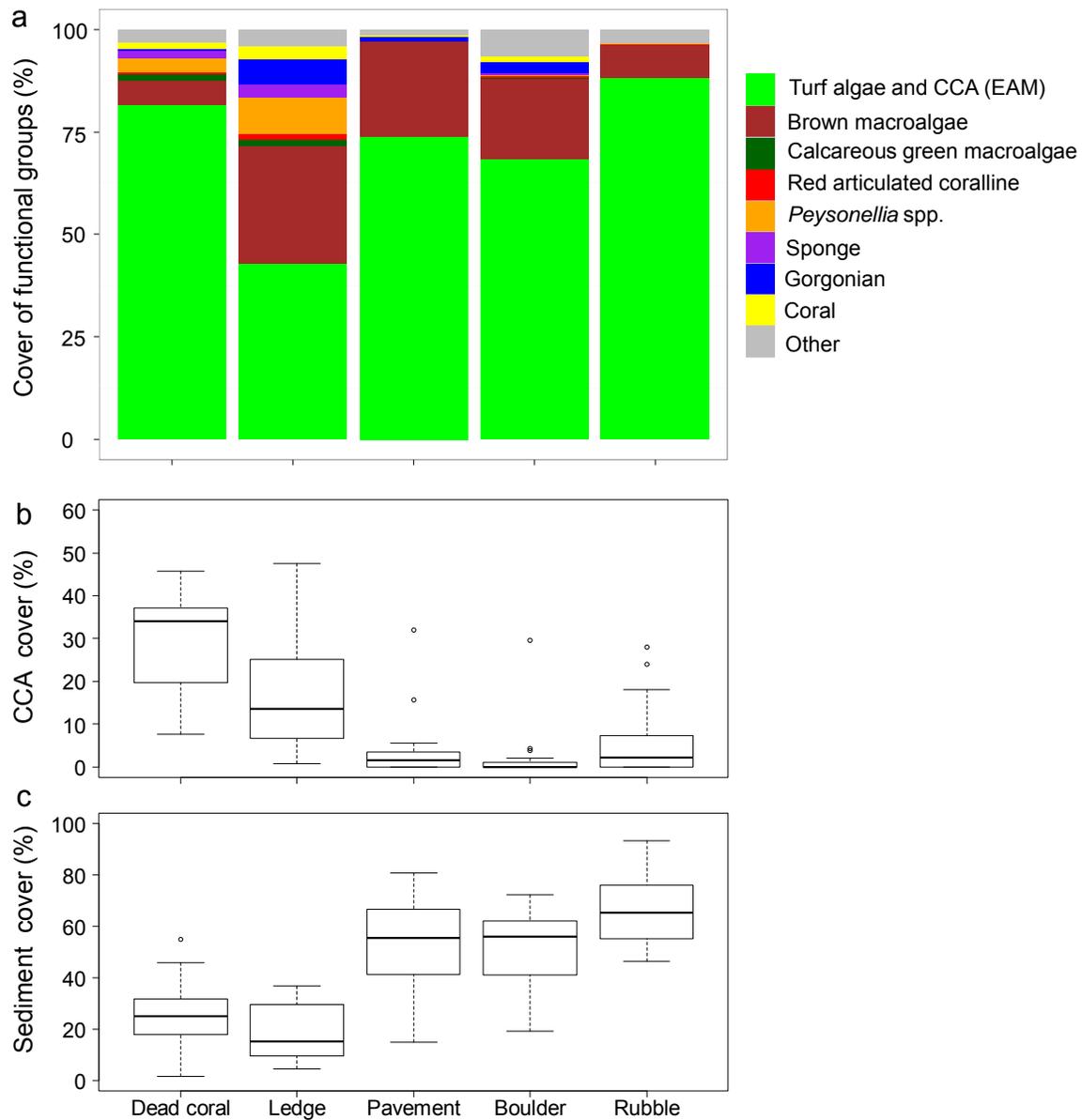


Figure S3. (a) Percent cover of benthic organisms on each of the five substrates commonly targeted by parrotfishes showing the relative availability of different food types (N = 24 per substrate). Note that food types are aggregated in order to match the resolution analyzed for the diet composition of parrotfishes. Box and whisker plots showing the percent of the ‘turf algae and CCA’ category that was covered in (b) CCA and (c) sediment. Note that dead coral and ledge substrates had relatively high cover of CCA and low levels of sediments, while pavement, boulder, and rubble substrates had low cover of CCA and high levels of sediments.

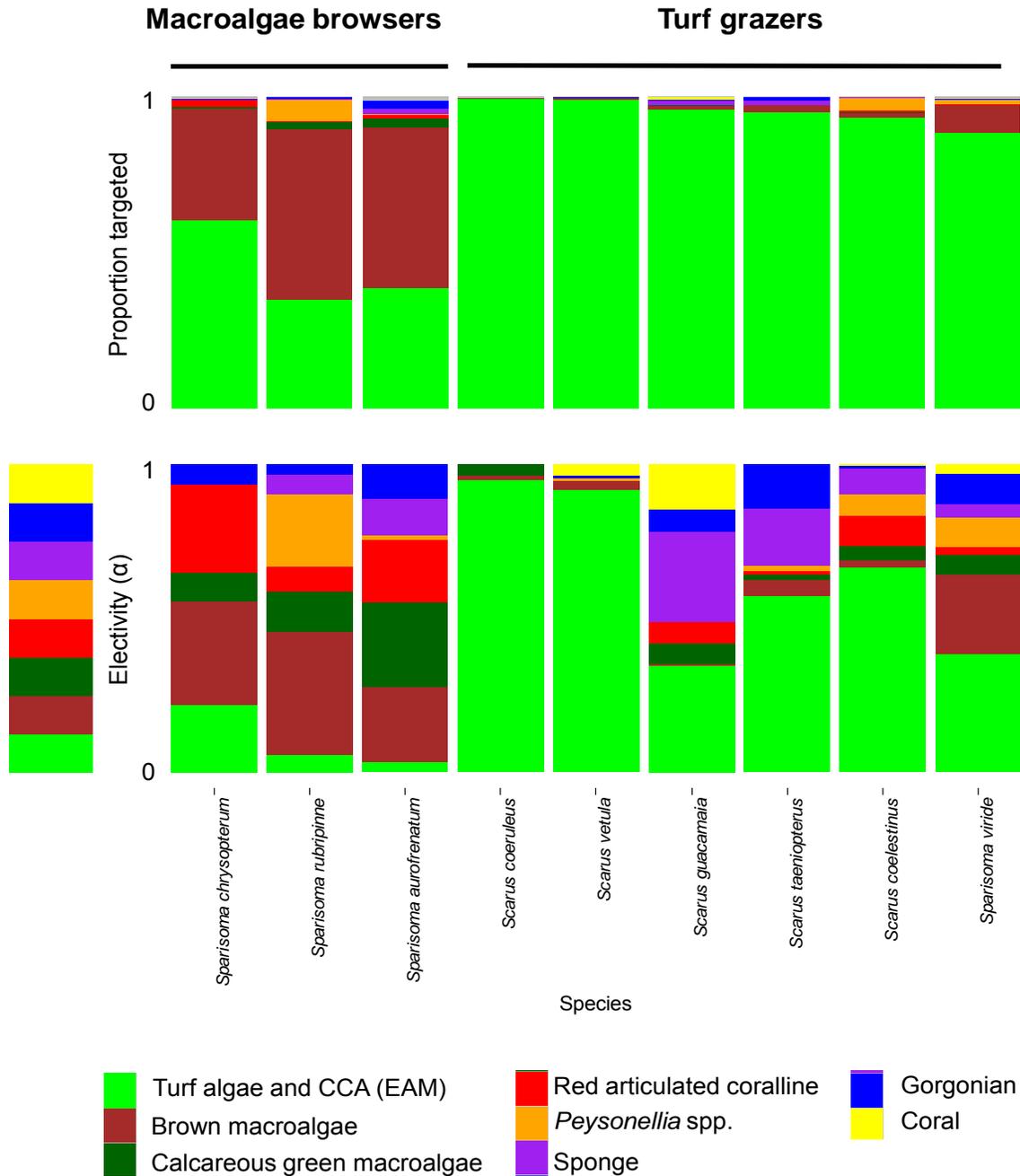


Figure S4. Diet composition of focal fish ( $N = \sim 18$  per species, see table S2 for exact sample sizes). The top row of bar charts shows the proportion of each food item targeted, and the second row shows electivity (Manly's alpha). The scale bar on the far left shows the expected pattern if all food items were targeted in proportion to their abundance (i.e., they were neither selected for or against). Electivity patterns reveal that many species also preferentially target less common food items, such as *Sp. chrysopterum* targeting red articulated coralline algae (*Amphiroa* spp., *Galaxaura* spp., *Jania* spp.), *Sp. aurofrenatum* targeting calcareous green algae (*Halimeda* spp.), and *Sc. guacamaia* targeting sponges and scleractinian corals. "Macroalgal browsers" and "Turf grazers" clusters are from species-averaged dendrogram (See Fig. 3).

Table S1. Sampling effort in different habitats

Site	Habitat Type	Area sampled (m <sup>2</sup> )
Carysfort	High-relief spur and groove	3408
	Low-relief hardbottom	2138
	Boulder/rubble fields	2545
Elbow	High-relief spur and groove	2804
	Low-relief hardbottom	685
	Boulder/rubble fields	3730
French	High-relief spur and groove	245
	Low-relief hardbottom	2314
	Boulder/rubble fields	1968
Molasses	High-relief spur and groove	3177
	Low-relief hardbottom	2533
	Boulder/rubble fields	3073

Table S2. Sampling effort for focal behavioral observations. Also note that the ~ 6 observations of each species at each site were approximately evenly distributed between 10:00 and 16:00

Species	Molasses reef	Carysfort reef	Elbow reef	Total
<i>Scarus coelestinus</i>	6	6	6	18
<i>Scarus coeruleus</i>	6	7	6	19
<i>Scarus guacamaia</i>	6	6	4	16
<i>Scarus taeniopterus</i>	7	6	6	19
<i>Scarus vetula</i>	5	6	6	17
<i>Sparisoma aurofrenatum</i>	6	6	6	18
<i>Sparisoma chrysopterum</i>	6	6	6	18
<i>Sparisoma rubripinne</i>	6	6	7	19
<i>Sparisoma viride</i>	6	6	6	18

Table S3. Loadings from PCA of food items targeted with the total variance explained by each principal component. Variables with the highest loadings on the first principal component (which explained 95% of the variation in diet) are indicated in bold.

	PC 1	PC 2	PC 3	PC 4	PC 5	PC 6	PC 7
Brown macroalgae	<b>-0.6597</b>	0.4859	0.4342	0.0710	-0.0508	0.0443	0.1483
Calcareous green algae	-0.0212	-0.0260	-0.4637	0.7799	-0.0548	0.1901	0.1648
Coral	0.0019	-0.0021	-0.0234	-0.0150	-0.0066	0.0114	-0.8965
Gorgonian	-0.0147	0.0021	-0.2476	-0.1346	-0.3369	-0.8278	0.1548
Red articulated corraline	-0.0144	-0.0423	-0.1723	-0.2172	0.8619	-0.0899	0.1831
Red encrusting	-0.0332	-0.7796	0.4971	0.0875	-0.0866	0.0219	0.1489
Sponge	-0.0030	-0.0253	-0.3354	-0.5543	-0.3590	0.5169	0.2112
Turf plus CCA	<b>0.7502</b>	0.3912	0.3813	0.0794	-0.0415	0.0294	0.1514
Variance explained	<b>0.9532</b>	0.0337	0.0057	0.0032	0.0020	0.0019	0.0003

Table S4. Loadings from PCA of substrates targeted with the total variance explained by each principal component. Variables with the highest loadings on the first and second principal components (which together explained 82% of the variation in substrates targeted) are indicated in bold.

	PC 1	PC 2	PC 3	PC 4	PC 5
Boulder	<b>0.6317</b>	0.0274	0.6051	0.2579	0.0293
Dead coral	<b>-0.5288</b>	<b>0.6806</b>	0.1294	0.2707	0.0217
Ledge	-0.0144	0.0077	0.0393	-0.6289	-0.6603
Pavement	<b>-0.4505</b>	<b>-0.7305</b>	0.1042	0.2925	-0.0186
Sand	0.0189	-0.0262	-0.1080	-0.5207	0.7413
Rubble	<b>0.3431</b>	0.0411	-0.7701	0.3286	-0.1134
Variance explained	<b>0.4613</b>	<b>0.3594</b>	0.1044	0.0427	0.0321

Table S5. Loadings from PCA of habitats selected with the total variance explained by each principal component. Variables with the highest loadings on the first principal component (which explained 74% of the variation in habitat electivities) are indicated in bold.

	PC 1	PC 2
Boulder and rubble	<b>0.7527</b>	0.3165
Low-relief	<b>-0.1022</b>	-0.8101
High-relief	<b>-0.6504</b>	0.4936
Variance explained	<b>0.7437</b>	0.2563

Table S6. Loadings from RDA with the total variance explained by each component

	RDA1	RDA2	RDA3	RDA4	RDA5	RDA6
Distance	-0.660	-0.514	0.604	0.419	-0.021	0.098
Bite rate	-0.854	1.608	-0.805	0.175	-0.055	0.053
Feeding_PC1	-1.042	1.452	0.773	-0.040	0.101	-0.097
Substrate_PC1	-1.029	-0.547	-0.265	-0.075	0.397	0.055
Substrate_PC2	0.440	0.778	0.403	-0.266	-0.029	0.182
Habitat_PC1	-2.110	-0.779	-0.032	-0.201	-0.221	0.007
Variance explained	0.472	0.383	0.109	0.019	0.013	0.003